

# MEMORANDUM

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## Environmental Effects of a Vehicle Tax Reform: Empirical Evidence from Norway

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# Environmental effects of a vehicle tax reform: empirical evidence from Norway\*

Alice Ciccone<sup>†</sup>

## Memo 03/2015-v1

(This version January 2015)

### Abstract

In 2007, the Norwegian government reformed the vehicle registration tax in order to reduce the carbon intensity of the new car fleet by incentivizing the purchase of more fuel efficient cars. This paper identifies the impact of the new tax structure on three main dimensions: (i) the average CO<sub>2</sub> emissions intensity of new registered vehicles, (ii) the relative change in sales between low and high polluting cars and (iii) the market share of diesel cars. A Difference in Difference approach is employed to estimate the short run effects on each outcome variable of interest. The results show that the average CO<sub>2</sub> intensity of new vehicles was reduced in the year of the implementation of the reform by about 7.5 g of CO<sub>2</sub>/km. This reduction is the result of a 12 percentage points drop in the share of highly polluting cars and of an increase of about 20 percentage points in the market share of diesel cars.

**Keywords:** CO<sub>2</sub> emissions intensity, New vehicles, Vehicle registration tax, Tax reform, Norway, Diesel.

**JEL:** H25, L62, Q51, Q53, Q54, R48.

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# 1 Introduction

Reducing automobile pollutant emissions is a critical step to mitigate climate change. The transportation sector, which produces 23% of world energy-related CO<sub>2</sub> emissions, is the second-largest sector of energy consumption. Almost three quarters of transport emissions come from road transport, specifically passenger cars and light-duty trucks.<sup>1</sup> There are many approaches to reducing transportation emissions including the development of more fuel-efficient vehicles; the use of alternative fuels that come from renewable resources such as biofuel, hydrogen, and electricity; and the reduction of demand for vehicle travel by, for example, improving public transportation or bike lanes. In practice, the most widely used economic incentives for reducing road transport emissions fall into two categories: "command and control" regulation, such as emissions standards, and market-based incentives, such as fiscal instruments like carbon taxes. These approaches are interconnected and complement each other. The European Commission, for instance, has set CO<sub>2</sub> emissions targets for manufacturers specifically directed at new passenger cars to improve fuel efficiency through technological development.<sup>2</sup> At the same time, EU-Member States are independently implementing diverse fiscal measures including vehicle taxes to encourage the purchase of new vehicles with lower CO<sub>2</sub> emissions, and fuel taxes or circulation taxes, to control transportation activities.<sup>3</sup>

This study assesses the effects of a recent policy intervention designed to influence the demand for private cars. In 2007, the Norwegian government reformed the Vehicle Registration Tax (VRT) system. The VRT is an upfront tax for new vehicles, which accounts for about half of the retail price. The government explicitly indicated that the objective of the reform was to reduce the intensity of CO<sub>2</sub> emissions of the average car fleet by incentivizing the purchase of more fuel-efficient cars.<sup>4</sup> The reform substituted the engine size component of the registration tax with the CO<sub>2</sub> emissions intensity component. This change increased the sensitivity of the tax to CO<sub>2</sub> emissions. As a consequence, consumers who purchase vehicles at the more efficient end of the distribution save about 10,000 NOK, while those who opt for relatively fuel-inefficient vehicles face an increase of about 50,000 NOK.

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<sup>1</sup>International Energy Agency [IEA \(2009\)](#) and IPCC report by [Kahn Ribeiro et al. \(2007\)](#).

<sup>2</sup>Targets are: 130 g of CO<sub>2</sub> per km for the average new car fleet by 2015 and 95 g of CO<sub>2</sub> per km by 2020. Regulation (EC) No 443/2009 and No 333/2014 of the European Parliament.

<sup>3</sup>See [van Essen \(2012\)](#) for an overview of carbon-based vehicles taxation schemes in the European Union.

<sup>4</sup>The CO<sub>2</sub> emissions intensity is a measure based on the expected grams of CO<sub>2</sub> that a vehicle will produce per kilometer driven and it is measured in gCO<sub>2</sub>/km.

What are the environmental effects of calculating the vehicle registration tax on CO<sub>2</sub> emissions intensity? To answer this question, I use a Difference in Differences (DID) approach to identify the short-run impact of the 2007 reform on three main outcome variables: 1) the average CO<sub>2</sub> emissions intensity of new registered vehicles, 2) the relative change between low and high polluting cars and 3) the market share of diesel cars.<sup>5</sup> This reduced-form approach offers a clear and simple identification of the response parameters of interest and is particularly well-suited for establishing causality (Timmins and Schlenker (2009)).<sup>6</sup> The choice of method is particularly appropriate because of the quasi-experimental nature of the phenomenon of interest. Other studies that have used reduced-form models to investigate related problems are Klier and Linn (2012), Klier and Linn (2013), Klier and Linn (2010), Hastings (2004) and Busse et al. (2006). The main data are provided by the Norwegian Road Federation OFVAS and contain repeated cross sections of new vehicles' monthly registrations in each municipality in Norway.<sup>7</sup>

The results suggest that the fiscal change induced a reduction of about 7.5 gCO<sub>2</sub>/km in the average CO<sub>2</sub> performance of new cars in 2007, which corresponds to 4.3% decrease from the pre-treatment average and account for about 20% of the standard deviation. Between 2006 and 2007 the average CO<sub>2</sub> intensity of the new car fleet dropped from 173 to 160 gCO<sub>2</sub>/km. The estimated causal impact of the reform, net of the anticipation effects, corresponds to about half of the overall reduction in the observed CO<sub>2</sub> intensity, which includes exogenous factors such as fuel efficiency improvements associated with the supply side of the market (Figure 1). The estimated reduction in CO<sub>2</sub> intensity is the combined result of a shift in demand toward greener vehicles and an increase in the market share of diesel cars. Specifically, the tax reform caused a reduction of about 12 percentage points in the share of high emitting vehicles, i.e. those emitting more than 180 gCO<sub>2</sub>/km, and an expansion between 19 and 21 percentage points of the share of diesel cars within the year of the reform. Furthermore, no significant change in the number of new sales is found indicating that the VRT reform has mainly induced substitution effects.

When using a fiscal instrument to reduce CO<sub>2</sub> emissions, it is critical to assess its effectiveness. Examples of such instruments include feebates, vehicle registration taxes, circulation taxes and fuel taxes. The use of these instruments for climate policies has become increasingly popular, but they have been implemented very differently by different countries. As a consequence, evidence of their economic and environmental effects is complex to assess and some-

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<sup>5</sup>The effects are estimated over a nine-month period within the year of the intervention.

<sup>6</sup> In this study, I aim to complement the large body of literature which makes use of structural models such as Bresnahan (1987); McCarthy (1996); Berry et al. (1995). Generally, these models do not focus primarily on vehicle taxes.

<sup>7</sup> Opplysningsrådet for Veitrafikken AS (OFV AS) <http://ofvas.no/>

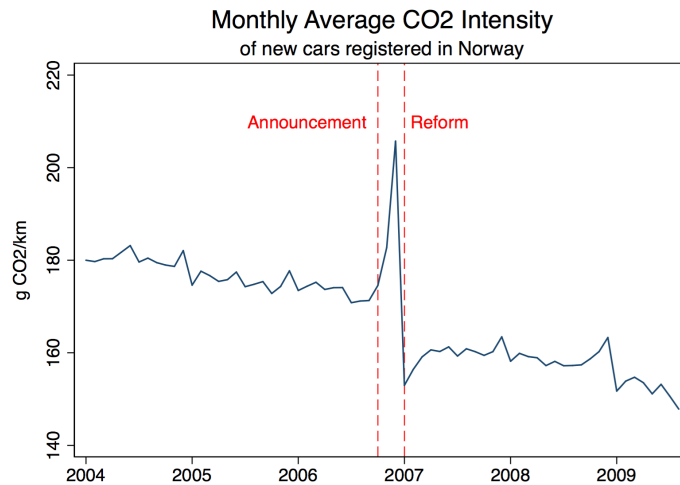


Figure 1: Monthly average CO<sub>2</sub> intensity of new vehicles registered in Norway between January 2004 and December 2011.

times conclusions are conflicting (Mandell (2009)). Some studies argue that upfront taxes, such as registration taxes, are the most effective instruments because they counterbalance consumer myopia in evaluating future costs. Consumers are more responsive to upfront prices and taxes than to the expected lifetime costs of the car (Allcott and Wozny (2012), Brand et al. (2013), Kågeson (2005), ICCT (2014), Greene et al. (2005)). However, registration taxes affect only new vehicles sales, whereas fuel and circulation taxes impact both new and used vehicles. Fuel taxes are effective as they act on two dimensions, discouraging the intensity of transport activities and encouraging the shift to more fuel efficient vehicles (Goodwin et al. (2004), Sterner (2007)). While there is disagreement about which policy is ultimately most effective, there is agreement that CO<sub>2</sub>-differentiated vehicles taxation can produce large reductions in emissions (COWI (2002)). Moreover, CO<sub>2</sub>-differentiated tax and feebate are more politically acceptable because of potential revenue neutrality, while an increase in fuel taxes is politically unpopular (Greene et al. (2005)).

The majority of studies on the effects of environmental policies are done ex-ante and are mostly based on simulations (BenDor and Ford (2006), Giblin and McNabola (2009), Greene et al. (2005)) and Skippon et al. (2012)). The present work belongs to the relatively small, but growing literature on ex-post evaluation of CO<sub>2</sub>-differentiated taxes which have been recently introduced in Europe. Various studies have shown that differentiating vehicle taxes on CO<sub>2</sub> emissions is an effective measure to reduce CO<sub>2</sub> intensity, though the magnitude of the results differs across countries and across instruments used. In 2008, France reformed its vehicle registration tax introducing a feebate

system.<sup>8</sup> As a consequence, an average decrease of 5% in CO<sub>2</sub> emissions is estimated in the short run, but the generosity of the subsidy increased the sales of new cars by 13% and cost 285 million euro to the state budget. Hence, [D’Haultfoeuille et al. \(2014\)](#) conclude that the environmental short-run impact of the feebate is negative. CO<sub>2</sub>-differentiated circulation taxes introduced in Germany and Sweden deliver less clear effects on the average CO<sub>2</sub> intensity of new vehicles compared with the French feebate, confirming that consumers are generally more responsive to upfront taxes ([Klier and Linn \(2012\)](#)). [Gerlagh et al. \(2015\)](#) consider 15 European countries and find that a one percent increase in the CO<sub>2</sub> sensitivity of registration taxes reduces the CO<sub>2</sub> intensity of the new fleet by 0.06-0.13 percent. The most relevant point of comparison for Norway is perhaps Ireland. Neither country has a car manufacturing industry, so their policies focus mainly on the demand side of the market. Like Norway, Ireland has substituted the engine size component of its vehicle registration tax with CO<sub>2</sub> performance. Ireland has gone a step further by also differentiating the circulation tax with respect to CO<sub>2</sub> emissions. Results from [Rogan et al. \(2011\)](#) are in line with the present work. In particular, they find a larger short term effect of about 13% reduction of CO<sub>2</sub> intensity, brought about primarily by the shift to diesel-powered vehicles.

The paper is organized as follows: background information specific to Norway is presented in the next section followed by a descriptive analysis of the Norwegian tax system for vehicles and the reform of the VRT in 2007. The data are presented in Section 3, where the main variables of interest are described. Section 4 explain in detail the empirical approach and the identification strategy for the outcome variables of interest. Section 5 discusses the main results, while section 6 presents effects on some additional variables. Finally, the conclusions are provided at the end of the paper.

## 2 Institutional Background

Oil was discovered in Norway in the 1960s, and as a result, Norway is one of the wealthiest countries in the world. The strict correlation between GDP and demand for private vehicles is generally well known, and Norway is no exception (Figure 18). Predictably, as the stock of private cars and mileage driven has increased, so have GHG emissions. Emissions of CO<sub>2</sub>, the main greenhouse gas, from Norwegian road transport are reaching levels almost 30% higher than those of 1990, making road transport one of the fastest growing sources of CO<sub>2</sub> emissions in the country.<sup>9</sup> However, Norway has

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<sup>8</sup>The feebate system consists in a subsidy for low-emitting vehicles and in a fee for cars emitting more than 160 g of CO<sub>2</sub>/km.

<sup>9</sup>Statistic Norway [www.ssb.no](http://www.ssb.no) (SSB: Statistisk Sentralbyrå)

taken preventive actions and thanks to the introduction of more energy-efficient vehicles as well as the blending of hybrids and electric cars, it is well in line with the goal of reducing CO<sub>2</sub> emissions intensity by 40% by 2020.<sup>10</sup> See Figure 19 for a graphical comparison of the CO<sub>2</sub> emission intensity trend for new vehicles in Norway and other European countries.

## 2.1 Vehicle Registration Tax

Purchase, ownership, and usage taxes serve as economic incentives to affect car purchase and usage decisions. In Norway, these policies are implemented through four elements. The registration tax (1) for new vehicles is a one-time fee paid at the moment of purchase and it accounts for almost half of the retail price. Ownership taxes for passenger cars consist of a flat annual circulation fee (2), and a reclassification fee (3), which applies only to used vehicles.<sup>11</sup> Fuel taxes (4) are determined by various factors including the CO<sub>2</sub> content of the fuel. Historically, the first three elements were primarily levied for state revenue, while fuel taxes reflect road use, accidents and other environmental costs.

This paper focuses on the vehicle registration tax (1). Since 1996, the registration tax has been proportionally linked to three characteristic of a vehicle: its weight; its engine size; and its power. In 2007, the component of the tax calculated according to engine displacement was substituted with the vehicle's potential CO<sub>2</sub> emissions intensity. In other words, beginning in January 2007, the registration tax on private vehicles became a stepwise function of weight (kg), power (kW), and CO<sub>2</sub> intensity (gCO<sub>2</sub>/km).

In order to understand the implications of the 2007 reform, it is important to assess the interdependence among the components of the tax, and how each component affects the monetary value of the registration tax. Engine size, power, and weight are all positively correlated with CO<sub>2</sub> emissions and respectively with each other (Table 1). Hence, by directly calculating the tax over CO<sub>2</sub> intensity as well as over weight and power, the total vehicle registration tax became more sensitive to CO<sub>2</sub> emissions than it was before the reform in 2007. The increase of CO<sub>2</sub> sensitivity is depicted in Figure 2, which highlights the relationship between the total registration tax in 2006 and in 2007 and the CO<sub>2</sub> intensity with linear fitted values differentiated by fuel. For levels between 200 and 300 gCO<sub>2</sub>/km the registration tax is higher in 2007 than it was in 2006. Moreover, the difference in tax between diesel and gasoline is reduced.

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<sup>10</sup>The reduction is compared with the level of 2007. Regulation (EC) No 443/2009 and No 333/2014 of the European Parliament.

<sup>11</sup>The ownership tax was differentiated by fuel type (gasoline or diesel) after 2008, hence it does not affect the present analysis.



Table 1: **Pearson correlation matrix**

	CO <sub>2</sub> int	Weight	Engine	Power	Diesel
CO <sub>2</sub> int	1				
Weight	0.6094	1			
Engine	0.6427	0.8312	1		
Power	0.6845	0.7378	0.8552	1	
Diesel	-0.1804	0.5188	0.3737	0.1558	1

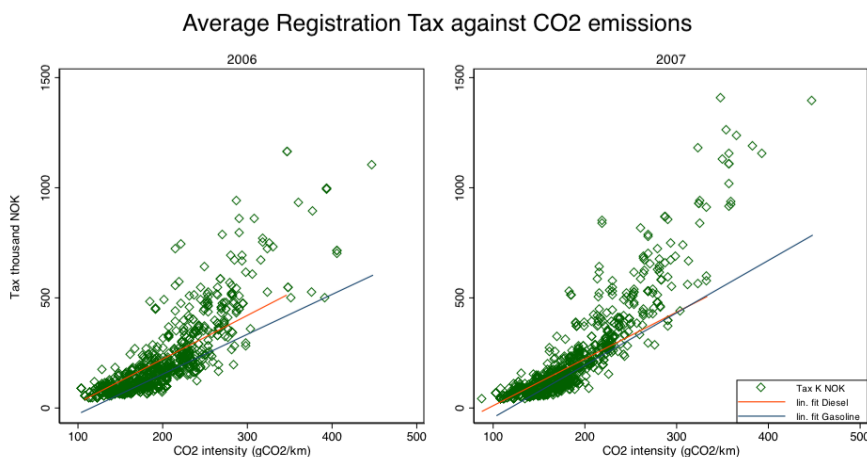


Figure 2: Scatter plot of average registration tax against CO<sub>2</sub> emissions intensity in 2006 and 2007. The two panels show the increase in CO<sub>2</sub> sensitivity of the registration tax before and after the reform. Linear fitted values for diesel and gasoline-fuelled vehicles show that the gap in tax between gasoline and diesel cars is reduced.

It is also possible to evaluate the change in tax paid before and after the reform of 2007 by clustering vehicles according to market segment. The first column of Table 2 illustrates the mean and standard deviation of the registration tax by market segments in 2006 together, with the most sold make and model. The second column shows the difference in tax in between 2006 and 2007. For mini, small, and compact cars, which are associated with lower emissions, weight, and power, the tax in 2007 was about 15% lower than in 2006. For example, buying a mini car, such as the Toyota Aygo, cost about 8 150 NOK, or 15 percent less in 2007 than in 2006. In contrast, consumers spent an average of 20% more to register larger cars, SUVs, or MPV. The registration fee for an SUV like the Suzuki Vitara increased on average by 25 500 NOK in 2007. More details and discussion regarding the registration tax is reported in the Appendix A.

Table 2: Average Change in VRT by Market Segment

Segment	Tax '06	Change '07
<b>Mini</b> (Toyota Aygo)	51.89 (3.47)	-8.15 (4.20)
<b>Small</b> (Toyota Yaris)	69.35 (9.81)	-11.35 (7.42)
<b>Compact</b> (Toyota Corolla)	103.77 (27.30)	-11.15 (26.75)
<b>Medium</b> (VW Passat)	141.81 (37.73)	-4.82 (23.07)
<b>SUV</b> (Suzuki Vitara)	216.17 (106.75)	+69.59 (85.10)
<b>MPV</b> (Ford S-Max)	191.96 (68.71)	+25.52 (32.45)
<b>Large</b> (Volvo V70)	241.58 (70.20)	+35.78 (85.30)

Thousand NOK (2012 currency). Standard Deviation in parenthesis.  
The make and model of the most-sold vehicle for each market segment is indicated in parenthesis.

### 3 Data

The main data used in this study were provided by the Norwegian Road Federation OFVAS<sup>12</sup> and contain detailed information about 670 000 new passenger cars sold in Norway between 2004 and 2009. I hold repeated cross section data with monthly registrations by vehicle specification in each municipality of Norway, i.e. panel data at month and municipality level.<sup>13</sup> Vehicle specifications are defined by brand, model, weight, engine displacement, power, potential CO<sub>2</sub> emissions, fuel type, number of doors, and transmission type. Vehicle-specific taxes have been calculated on the basis of these characteristics following the scheme provided by OFVAS.

Other information regarding population, yearly gross income per capita for Norwegian municipalities, and average fleet age at the county level was provided by Statistic Norway (SSB).<sup>14</sup> Monthly average fuel prices and fuel taxes for both gasoline and diesel in Norway were provided by the Institute of Transport Economics (TØI).<sup>15</sup> Summary statistics for the most relevant vehicle characteristics are reported in Table 3.

<sup>12</sup> Opplysningsrådet for Veitrafikken AS (OFV AS) <http://ofvas.no/>

<sup>13</sup> Norway counts 428 municipalities (kommuner) in 2013.

<sup>14</sup> Statistisk Sentralbyrå, [www.ssb.no](http://www.ssb.no).

<sup>15</sup> [www.toi.no](http://www.toi.no)

Table 3: Most sold model, total number of new vehicles registered each year and mean of the main vehicles' characteristics.

Year	Top sold model	Tot cars sold	CO <sub>2</sub> int. (g/km)	Weight (Kg)	Power (kW)	Diesel Share
2004	Toyota Av.	115600	180.44	1332.76	85.29	28%
2005	Toyota Cor.	109846	175.62	1349.55	85.23	39%
2006	VW Passat	109098	177.32	1402.94	89.92	48%
2007	VW Passat	129121	159.40	1403.05	86.97	74%
2008	VW Golf	110540	158.99	1418.45	89.78	73%
2009	VW Golf	98640	151.35	1413.07	89.6	73%

Source: [www.ofvas.no](http://www.ofvas.no)

### 3.1 Main variables

This study identifies the causal effects of 2007 reform of the registration tax on three main outcome variables: CO<sub>2</sub> emissions intensity, the share of high polluting cars, and the market share of diesel cars. The following descriptive analysis reveals several important changes in the characteristics of the passenger vehicle fleet in Norway between 2004 and 2009. The overall decline of the trend illustrated in Figure 1 is due in part to improvements in fuel efficiency of the vehicles available on the market and in part to a shift on the demand side. Demand-side responses by Norwegian consumers include a shift toward less polluting cars and toward diesel-fuelled, rather than gasoline-fuelled, vehicles. The next sections describe in more details each of the three main outcome variables of interest.

#### CO<sub>2</sub> emissions intensity

The average CO<sub>2</sub> intensity of the new car fleet decreased by almost 30 gCO<sub>2</sub>/km: from an average of 180 in 2004 to about 151 gCO<sub>2</sub>/km in 2009 (Figure 1 and Table 3). In October 2006, the Norwegian Ministry of Finance presented a proposal for the 2007 national budget which included the suggested change to the vehicle registration tax system.<sup>16</sup> Public discussion of the proposed changes led to a high level of anticipation behavior in November and December 2006. This announcement effect is identifiable in Figure 1, where the drastic increase in average CO<sub>2</sub> intensity in November and December 2006 is followed by a drop in CO<sub>2</sub> intensity beginning in January 2007 when the reform was activated. The density of the average CO<sub>2</sub>

<sup>16</sup>Stortingsproposisjon nummer 1 (2006-2007) <http://www.statsbudsjettet.no/Statsbudsjettet-2007/>

intensity shifted toward lower emissions in 2007 compared with 2006 (Figure 3). Most of the distribution became concentrated below the 180gCO<sub>2</sub>/km cutoff. Figure 3 also shows some threshold effects. Specifically, there is a reduction right after the 140 and the 180 cut off which was absent in 2006.

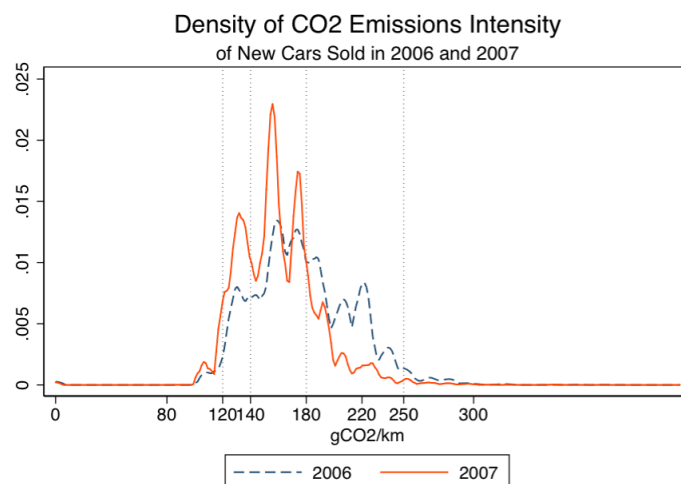


Figure 3: Density of CO<sub>2</sub> emissions intensity in the years before and after the reform. Vehicles with zero emissions are electric, gas, hybrid and hydrogen cars which account for about 2% of the sample.

### Low- versus high-polluting vehicles

Examining the purchases of different classes of vehicles is another way to understand what happened in the months immediately before and after the 2007 reform. Figure 4 shows the market share of new cars purchased by the CO<sub>2</sub> thresholds used in the calculation of the registration tax. It appears that the discontinuity observed in the CO<sub>2</sub> intensity trend is an intertemporal substitution between high- and low-polluting cars. This substitution is clearly visible in Figure 4, where opposite trends are depicted for different kinds of vehicles in the two panels. Sales of vehicles with low CO<sub>2</sub> intensity in the left panel, i.e. those emitting less than 180 gCO<sub>2</sub>/km decrease before January 2007 and increase immediately after. Vehicles with high CO<sub>2</sub> intensity reported in the right panel show the opposite trend. Figure 5 plots only the share of highly-polluting cars, which is the second outcome variable considered in this study. The share of cars emitting more than 180 gCO<sub>2</sub>/km decreased by almost 30 percentage points, from an average of 43% in 2004 to about 16% in 2009.

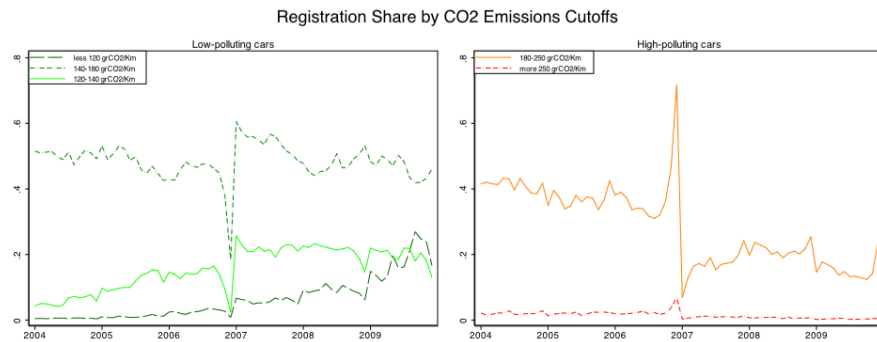


Figure 4: Share of new vehicles registered by CO<sub>2</sub> intensity category. The categories are made by taking into account how the CO<sub>2</sub> component of the tax is structured. Note that the two panels show market shares so they sum to one.

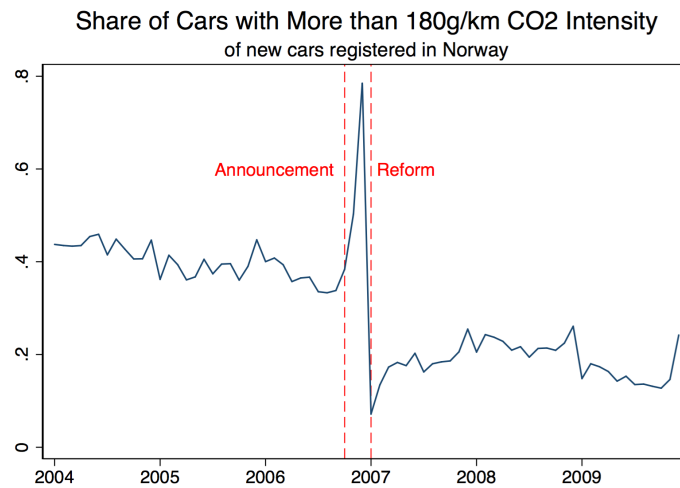


Figure 5: Market share of new registered vehicles with more than 180 g per km of CO<sub>2</sub> intensity.

### Diesel-powered vehicles

The market share of diesel cars increased greatly between 2004 and 2009 (Figure 6). From levels around 28% in 2004, diesel-powered vehicles reached levels around 73% in 2009 (Table 3). The increasing trend can be partially explained by specific taste for the superior fuel efficiency of diesel engines and the relatively lower price of fuel. In Norway, fuel prices are lower for diesel than for gasoline (Figure 20), but vehicle taxes favor gasoline cars (Figure 12). Because diesel cars are generally associated with lower CO<sub>2</sub> intensity, after the 2007 reform, diesel cars became relatively cheaper to buy (Figure

12). Specifically, the difference in registration tax for diesel and gasoline cars decreased from an average of 56 000 NOK in 2006 to an average of 32 000 NOK in 2007. This paper demonstrates that part of the increase in the market share of diesel cars was the result of the registration tax reform of 2007 and this pattern is consistent with other studies such as Gerlagh et al. (2015) and Rogan et al. (2011). More discussion regarding the characteristics of diesel cars is reported in the Appendix A.

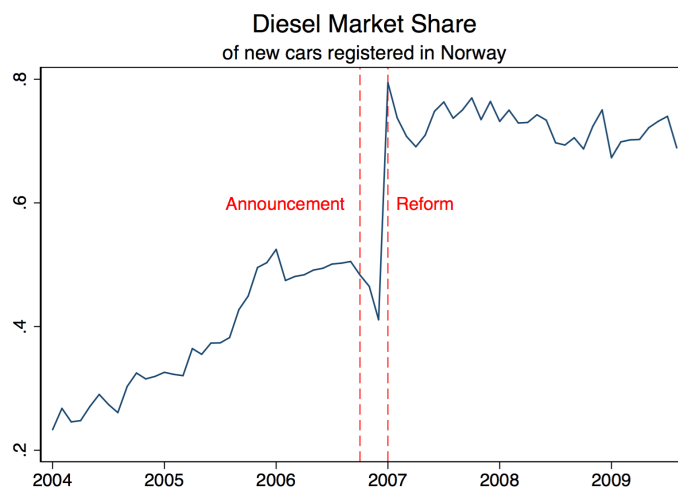


Figure 6: Diesel share from January 2004 until December 2009

## 4 Empirical Approach and Identification.

What is the impact of differentiating the vehicle registration tax on CO<sub>2</sub> emissions intensity? To answer this question, I use a Difference in Differences (DID) approach and exploit the reform of 2007 to estimate the causal environmental effects, in the short run. Specifically I estimate the impact of the reform on three outcome variables: average CO<sub>2</sub> emissions intensity, the share of highly polluting cars, and the share of diesel vehicles. In this section I explain how much of the changes observed in these variables are associated with the tax reform of 2007.

By using the DID estimator it is possible to calculate the causal effect net of time trends and market seasonality, and control for exogenous factors that are potentially relevant for the outcome variables. For instance, the average CO<sub>2</sub> intensity of vehicles purchased in the nine months before the reform is about 12.6 g of CO<sub>2</sub> per kilometer higher than the average for those bought after the reform. Considering only this simple difference, however, produces

a biased estimation of the real change in the average CO<sub>2</sub> intensity of the new fleet as long as the time trend is non-zero. As discussed previously, the reduction in CO<sub>2</sub> intensity is in fact due to both a supply and a demand effect.

The DID method is used to evaluate the impact of a treatment on an outcome variable over a population. Generally, the population is divided in two groups: those who receive the treatment (the treated), and those who do not (the control group). This allows for a direct comparison, under specific assumptions, between the control and the treatment group. In this paper, the tax reform was applied to all vehicles in the market at the same time in Norway. Hence, there is no optimal control group in the standard sense. When dealing with cross sectional data, however, it is possible to compensate for the lack of control group by employing previous observations in time, when comparable to the primary observations of interest. Similar strategies have been used by [Schönberg and Ludsteck \(2012\)](#), [Lalive and Zweimüller \(2009\)](#), [Lalive et al. \(2010\)](#) and [Ekberg et al. \(2013\)](#).

In order to isolate the causal impact of the reform, observation of the outcome variables in previous years, when no reform took place, are used as a control. Specifically, two nine-month periods in 2006 and 2007 are used as the treatment observations, one nine-months period before the intervention and one period after. Two corresponding nine-month periods in 2004 and 2005 are used as control observations. A visual comparison of treatment and control is presented in [Appendix B](#).

As indicated in the previous section, the registration tax reform was announced in October 2006. From [Figure 1](#) it is clear that the announcement of the reform led to a high level of anticipation behavior where highly polluting vehicles experienced an extraordinary increase in their purchase before January 2007. Threats to identification can arise when individuals change their behavior as a consequence of the treatment, or in anticipation of it. Therefore, the months between October 2006 and March 2007 are excluded from the analysis. [Appendix C](#) reports robustness checks where the after-treatment period is postponed to check whether the adjustment period was longer than three months.

Following a standard DID procedure, Equation (1) is estimated for three outcome variables ( $Y_{r,t}$ ): the average CO<sub>2</sub> emissions intensity, the share of high polluting cars, and the share of diesel cars bought. The level of aggregation used to calculate the averages is municipalities  $r = 1, 2, \dots, R$

and months  $t = \{t_1, t_2, t_3, t_4\}$ .<sup>17</sup>

$$\bar{Y}_{r,t} = \alpha + \beta Reform_{t_3,t_4} + \gamma After_{t_2,t_4} + \delta After \cdot Reform_{t_4} + \mu C'_{r,t} + \epsilon_{r,t} \quad (1)$$

Where

- $t_1$  identifies the months between January and September 2004,
- $t_2$  is equal to one for observations between April and December 2005,
- $t_3$  identifies the months between January and September 2006, and
- $t_4$  identifies the months from April to December 2007.

The regressor  $Reform_t = 1$  for  $t = \{t_3, t_4\}$  is a dummy variable indicating which observations belong to the years of the treatment, namely  $t_3$  and  $t_4$ . The variable  $After_t = 1$  for  $t = \{t_2, t_4\}$  identifies the periods after the treatment in the year of the reform and for the control group. The variable takes the value 1 for the months between April and December 2005 and from April to December 2007 and zero otherwise.  $After_t \cdot Reform_t = 1$  for  $t = \{t_4\}$  is the interaction term identifying the nine-month period after the treatment. Lastly,  $C'_{r,t}$  is a vector of control variables.  $\epsilon_{r,t}$  is a random, unobserved error term.

In order to have an unbiased estimation of the treatment effect  $\delta$ , some assumptions must be verified. The model needs to be correctly specified and the error term needs to be uncorrelated with the variables in the equation. The identifying assumption is that treatment and control observations differ only because of the treatment. In other words, the Common Trend Assumption (CTA) states that in absence of intervention treatment and control groups would have common trends. This assumption is, in principle, untestable, however, a testable implication is that the pre-intervention trends in the control and treatment should be parallel. This implication is often used to infer the plausibility of the CTA. This study compares the trends for the time periods employed in the estimations in Appendix B.

The similarity in trends, lies in the seasonality of the car market. The comparability between treatment and control observations is reasonable given the regularity in the production cycles for cars. In the European market, one cycle correspond to a calendar year, meaning vehicles characteristics are constant for twelve-month periods. It is enough to attribute the observable trend in the outcome variables as an exogenous factor intrinsic with the supply side, and exploit the same months from the previous year as a feasible control group. To visually verify the seasonal regularity of the car market, the treatment group and control group are compared in Appendix B. If we

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<sup>17</sup>The total number of municipalities included in the analysis is 437 because some municipalities were split and other joint during the years considered in the study. As a consequence the panel is not perfectly balanced.



exclude the months immediately before and after the reform, the trends in the years of interest are analogous. This guarantees that the causal effect of the reform can be identified as the gap between the trends before and after the reform.

Possible threats to identification can arise in presence of exogenous factors affecting the outcome variables differently in different years/months. This study includes relevant time-varying control variables: annual per capita gross income at the municipality level, monthly fuel prices in Norway, and the average fleet age at the county level.

## 5 Results

This section presents and discusses the overall findings of the paper. To evaluate the causal impact of the registration tax reform of 2007, the treatment effect is estimated for each outcome variable of interest  $Y_{r,t}$  using Equation 1. The averages for the outcome variables are calculated at the municipality level for the 36 months used in the analysis.

The treatment effect  $\delta$  can also be calculated in a more direct and intuitive way by following the definition of the DID estimator. See, for example, Table 4 which exemplifies this calculation for the outcome variable CO<sub>2</sub> intensity. The DID estimator calculates the difference between the pre- and post-reform averages in the year of the intervention (treatment) minus the difference between the same time intervals in the control.

Table 4: **Treatment effect for CO<sub>2</sub> intensity calculated by applying the DID definition.**

CO <sub>2</sub>	Post-Reform	Pre-Reform	Diff
Treatment	$\bar{Y}_{r,t_4} = 160.57$	$\bar{Y}_{r,t_3} = 173.12$	-12.55
Control	$\bar{Y}_{r,t_2} = 175.38$	$\bar{Y}_{r,t_1} = 180.53$	-5.15
<b>Diff</b>	-14.81	-7.41	<b>-7.4</b>

Comparing the CO<sub>2</sub> intensity averages before and after the reform, we find a reduction of 12.55 g of CO<sub>2</sub> per km. This result cannot be interpreted as a consequence of the intervention. The observed reduction is due to a combination of improvements in fuel efficiency of the vehicles available on the market and a shift in the demand side, which could be a reaction to the 2007 reform of the registration tax. Using the DID approach we learn that the causal impact of tax reform on the demand is about 60% of the overall reduction observed before and after the reform. Specifically, Table 4 reports a reduction of 7.4 gCO<sub>2</sub>/km, which corresponds to about a 4.3% of the pre-treatment average. This simple calculation can be compared with the results of the OLS estimation reported in Table 5. The advantage of

OSL is the possibility of introducing control variables and the convenience of calculating standard errors.

All models reported in Table 5 are weighted on the number of car sold and have robust standard errors clustered on municipalities to account for similarities in demand in different time periods within the same municipality. For each outcome variable of interest Column (2) includes potentially relevant time-varying control variables: gross income at municipality level, the ratio between diesel and gasoline fuel prices, and the average age of the car fleet in the 19 Norwegian counties. All the treatment effects are highlighted in the first row of Table 5. They are highly significant and stable when including control variables.

For the outcome variable CO<sub>2</sub> intensity, the model estimates a change between 7.4 and 7.6 gCO<sub>2</sub> per km. The interpretation of these coefficients is that the reform caused a reduction in average CO<sub>2</sub> intensity of about 4.3% from the pre-treatment average within 2007 net of time trends and anticipation effects. This effect is quite large since it accounts for about 20% of the overall standard deviation. The estimated causal impact of the reform corresponds to about 60% of the overall reduction in the CO<sub>2</sub> intensity observed between 2006 and 2007, which includes exogenous factors such as fuel efficiency improvements associate with the supply side of the market.

To understand the mechanism behind the estimated reduction of CO<sub>2</sub> emissions intensity of the car fleet, this study also considers the impact of the tax reform on the share of highly polluting vehicles relative to medium-low polluting vehicles and on the share of diesel cars. Demand-side responses by the Norwegian consumers include a shift toward less polluting cars and toward more diesel-fuelled cars. Table 5 shows that the reform caused a decrease in the share of highly polluting vehicles (those emitting more than 180g of CO<sub>2</sub> per km) of about 12 percentage points. Diesel cars, instead, increased their market share between 19 and 21 percentage points in the short run. Figure 5 shows that the share of cars emitting more than 180 g CO<sub>2</sub>/km decreased by almost 30 percentage points, from an average of 43% in 2004 to about 16% in 2009. Hence, the reform of 2007 caused almost half of the observed reduction within the same year. At the same time we observed a general increase in the share of diesel, from levels around 28% in 2004 diesel-powered vehicles reach levels around 73% in 2009 (Figure 6). The increase in diesel share can be attributed to the substitution of the CO<sub>2</sub> intensity component for the engine size component of the registration tax. Given the higher engine size and the lower CO<sub>2</sub> emissions associated with diesel vehicles, the overall tax makes diesel cars relatively cheaper in 2007 than in 2006 (Figure 12). The remaining part of the trend is associated with a specific taste for diesel powered vehicles which have, on average, a much higher fuel economy than gasoline cars and run on a cheaper fuel (Figure

20).

Table 5: Estimation results

	(1)	(2)	(1)	(2)	(1)	(2)
	CO2 int.	CO2 int.	High-poll.	High-poll.	Diesel	Diesel
<b>Treatment eff</b>	-7.398*** (0.491)	-7.608*** (0.444)	-0.124*** (0.00662)	-0.116*** (0.00675)	0.193*** (0.00824)	0.207*** (0.00786)
Group effect	-7.414*** (0.314)	-7.845*** (0.318)	-0.0731*** (0.00389)	-0.0823*** (0.00401)	0.175*** (0.00598)	0.176*** (0.00735)
Time effect	-5.153*** (0.292)	-6.198*** (0.507)	-0.0493*** (0.00545)	-0.0662*** (0.00795)	0.104*** (0.00521)	0.111*** (0.00698)
Income		0.0407*** (0.0106)		0.000450*** (0.000118)		-0.000527*** (0.000149)
Diesel/Gas price		-2.426 (2.540)		0.121*** (0.0329)		0.203*** (0.0286)
Fleet age		-0.118 (0.522)		-0.00161 (0.00644)		0.0259*** (0.00558)
Constant	180.5*** (0.911)	172.0*** (6.606)	0.438*** (0.0107)	0.216* (0.0841)	0.0876*** (0.00643)	-0.198* (0.0963)
Observations	13813	13577	13813	13577	23833	23430
Adjusted $R^2$	0.497	0.534	0.429	0.455	0.307	0.331

Robust standard errors clustered on municipalities.

## 6 Other Variables

This section reports the impact of the 2007 reform on other characteristics of the vehicles. It is plausible to expect some change in other characteristics such as weight, power, and engine size. Moreover, given the increase of the share of diesel-fuelled cars, it is important to consider whether these vehicles have characteristics that differentiate them from the average fleet. As Figure 11 shows, the engine size, weight, and power of diesel cars are higher than those of gasoline cars. Figure 7, 8, and 9 show the trends for weight, power, and engine size. There seems to be a small increase for weight and power, but it is not clear whether the reform created any effect in the short run.

Table 6 reports the estimations for the outcome variables weight, power, and engine size. There is no significant increase in weight and a low significant increase for power. In contrast, the average engine size increased significantly in the short run. The estimated effect is about 47 ccm more in 2007. This effect correspond to a 2.6% increase from pre-treatment averages. However, from Figure 9 we notice a decrease in the trend after 2008, so even if the reform has created a short run effect, this effect disappears in the longer run.

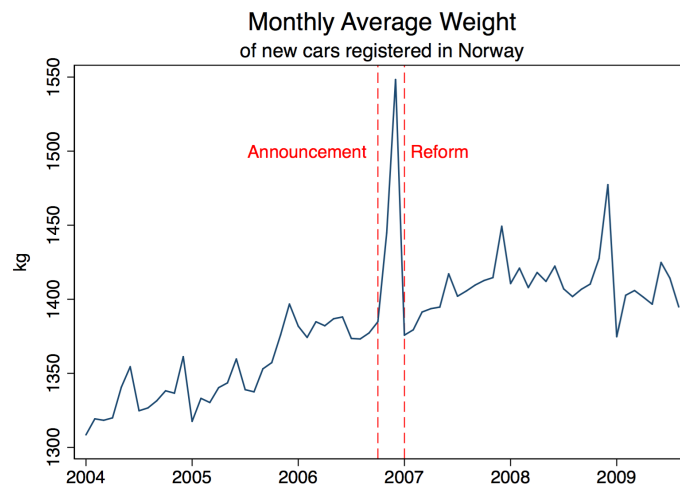


Figure 7: The average weight of new registered vehicles increases until 2008 and then stabilises afterward.

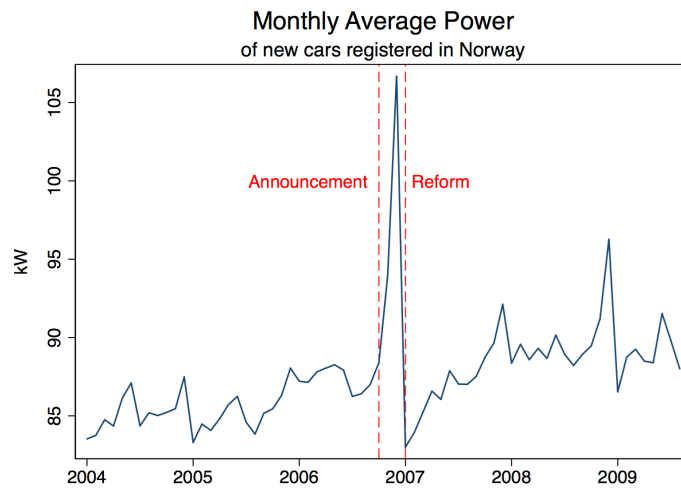


Figure 8: The average power of new registered vehicles increases slightly with time.

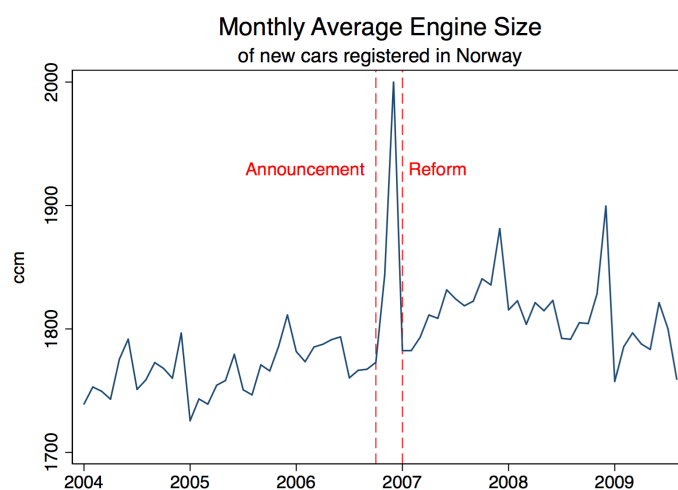


Figure 9: The average engine size of new registered vehicles increases slightly until 2008 and then decreases.

Lastly, it is also important to wonder whether the reform has induced an increase or decrease in the total number of cars sold. Figure 18 shows how the total number of sales are correlated with GDP per capita. The first column of Table 6 reports the estimated effect of the reform on the total number of registrations. The coefficient is non-significant, hence we cannot attribute any change in the number of cars sold to the implementation of the reform.

Table 6: Estimation results of additional variables

	Weight	Power	Engine size	Registrations
<b>Treatment eff</b>	6.064 (3.789)	0.569* (0.282)	46.71*** (4.291)	-1.229 (1.532)
Group effect	48.94*** (2.500)	1.744*** (0.220)	13.30*** (3.804)	-15.35*** (4.576)
Time effect	20.94*** (4.370)	-0.569 (0.448)	-1.979 (6.012)	-17.66*** (5.082)
Gross Income	0.191* (0.0933)	0.0332** (0.0114)	0.324* (0.159)	0.862*** (0.257)
Diesel/Gas price	74.56*** (21.00)	9.708*** (2.345)	88.70* (35.57)	-22.73*** (4.419)
Fleet age	7.439 (4.657)	0.0757 (0.550)	8.936 (8.045)	-3.104 (4.883)
Constant	1131.0*** (57.42)	65.94*** (6.870)	1496.8*** (106.9)	-136.9 (71.25)
Observations	13577	13577	13577	13577
Adjusted $R^2$	0.251	0.124	0.111	0.150

Robust standard errors clustered on municipalities.

## 7 Conclusions

This paper analyses how a recent policy intervention has affected the main characteristics of the new car fleet in Norway. In 2007, the structure of the vehicle registration tax changed. Taxation for new passenger cars became based on expected CO<sub>2</sub> emissions per kilometer rather than engine size. This reform was implemented with the goal of reducing the average CO<sub>2</sub> emissions intensity of the fleet. Using observational data on car purchases provided by the Norwegian Road Federation OFVAS, I estimate with a Difference in Differences approach the short run effect of the 2007 reform on three dimensions: 1) the average CO<sub>2</sub> intensity of new registered vehicles, 2) the relative change between low and high polluting cars in the market and 3) the market share of diesel cars.

The change in the tax structure did indeed result in an important reduction of the average CO<sub>2</sub> intensity. The estimated treatment effect is about 7.5 g of CO<sub>2</sub> per km less than we would have had without the reform. This reduction accounts for about half of the overall reduction in CO<sub>2</sub> intensity when including exogenous fuel efficiency improvements associated with the supply side of the market. The observed improvement in CO<sub>2</sub> performance is the result of a shift in demand toward greener vehicles and an increase of the diesel market share. The reported results show that, within the year of the policy implementation, the share of highly polluting vehicles dropped by about 12 percentage points as a result of the policy reform. Further, because of the substitution between the engine size component and the CO<sub>2</sub> intensity component, the market share of diesel cars has increased in the short run by about 20 percentage points.

This study also investigates whether the implementation of the reform has affected other characteristics such as weight, power, and engine size of the new purchased vehicles and whether it affected the total number of registrations. No significant change is estimated for the number of registration and for weight. A small and weakly significant increase is found in the average power of new cars. Lastly, a significant increase of about 2.6% in the average size of the engines is estimated.

## Appendix

### A Vehicle Registration tax

Table 7 shows the tax bands for each element used to calculate the registration tax for the years considered in this analysis. Before 2007, the VRT was calculated using three characteristic of a vehicle: its weigh (kg); its engine displacement (cylinder capacity cm<sup>3</sup>); and its power (kW). From January 2007, it was calculated based on the weight (kg), power (kW), and CO<sub>2</sub> intensity (g/km) of the purchased vehicle.

Table 7: Bands for the VRT components in different years

		2004	2005	2006	2007	2008
<b>Weight (kg)</b>	0-1150	39.52	39.76	39.16	36.82	36.40
	1151-1400	79.04	79.52	79.45	80.25	79.32
	1401-1500	158.10	159.05	157.77	160.52	158.67
	over 1500	183.87	184.97	183.51	186.68	184.53
<b>Power (kW)</b>	0-65	152.66	153.58	153.30	133.91	132.37
	66-90	556.79	560.14	557.24	557.97	551.55
	91-130	1113.93	1120.63	1115.59	1339.12	1323.71
	over 130	1885.04	1896.37	1886.54	2789.83	2757.73
<b>Engine Vol (ccm)</b>	0-1200	11.67	11.74	11.68		
	1201-1800	30.55	30.73	30.58		
	1801-2200	71.86	72.29	71.94		
	over 2200	89.77	90.31	90.42		
<b>gCO<sub>2</sub>/km</b>	0-120				44.64	44.13
	121-140				212.03	209.59
	141-180				557.97	551.55
	181-250				1562.30	1544.54
	over 250				1562.30	1544.54

Prices are in NOK (2012 currency)

Before 2007, the weight component accounted on average for 54% of the overall registration tax, while the power and engine size components accounted for 19% and 27% respectively. After the intervention, the new CO<sub>2</sub> intensity component accounted for about 18% of the total tax, while the power component remained quite stable around 20%. As a consequence, the vehicle's weight became more important and accounts for about 60% of the overall registration tax (Figure 10).

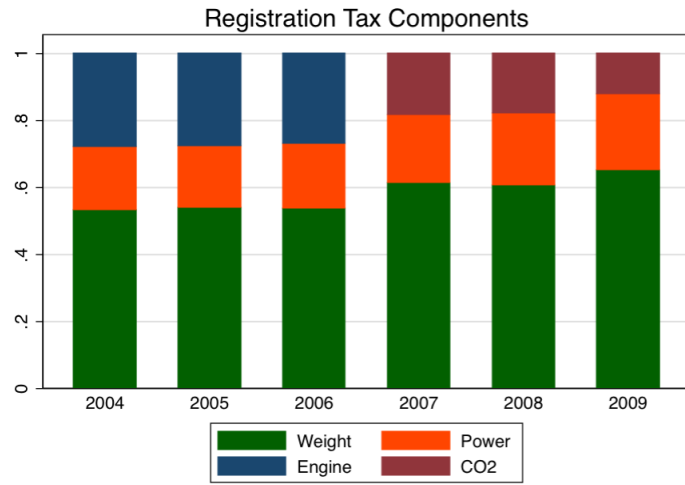


Figure 10: Vehicle registration tax components and their weights over the years.

To assess the effect of the reform on the tax it is useful to separate vehicles by fuel type. Diesel cars have, on average, larger engine size, higher weight, and higher power, but lower CO<sub>2</sub> intensity than their gasoline equivalents (Figure 11).

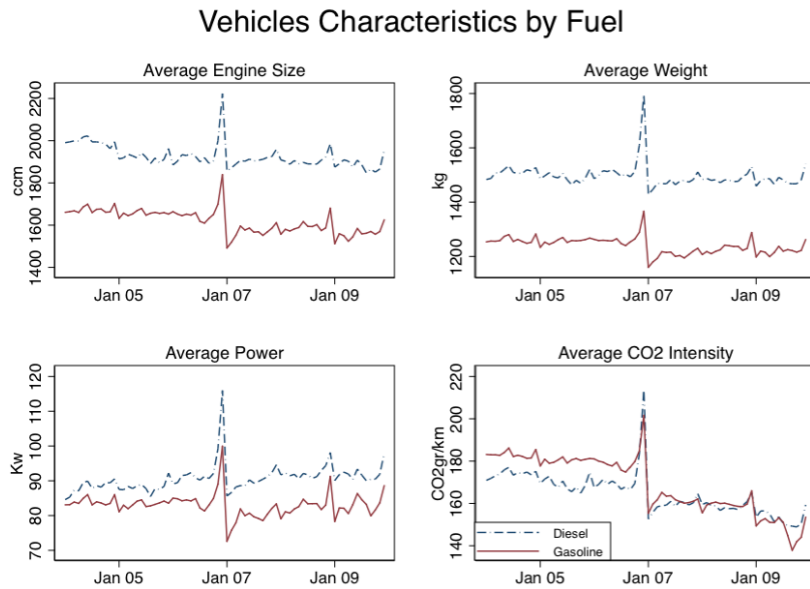


Figure 11: Average engine size, weight, power, and CO<sub>2</sub> intensity of new registered vehicles by fuel type.



This meant that when the registration tax was calculated based on the engine size, diesel-fuelled vehicles were more expensive than gasoline cars with similar characteristics. Because of the substitution of the engine size component with the CO<sub>2</sub> emissions intensity, diesel cars became relatively cheaper (Figure 12). Specifically, the difference in registration tax for diesel and gasoline cars decreased from an average of 56 000 NOK in 2006 to an average of 32 000 NOK in 2007.

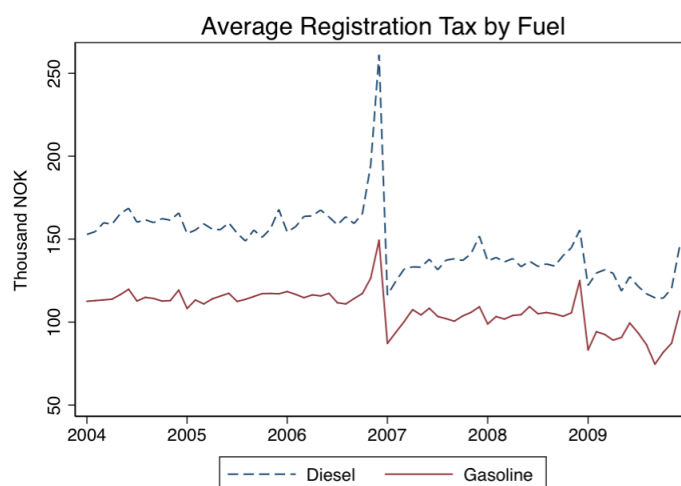


Figure 12: Average vehicle registration tax by fuel type. Diesel vehicles pay a higher registration tax than gasoline cars. However, after the 2007 reform the price gap is reduced.

By plotting each component of the registration tax against the total value of the VRT expressed in thousand NOK, it is possible to understand how the tax has changed over time, and in particular, how the total value of the tax changed after the reform in 2007. The marginal tax rate for the weight component is almost unchanged through the years, while the marginal rate based on the car's power increased in 2007, making vehicles with engine power higher than 130 kW more expensive (Figure 13).

Comparing the CO<sub>2</sub> intensity component introduced in 2007 with the engine size component of 2006, it is clear that the tax calculated over the CO<sub>2</sub> emissions factor is steeper (Figure 14). As a consequence, the monetary value of registration tax became higher for highly-polluting vehicles compared to tax values registered in 2006.

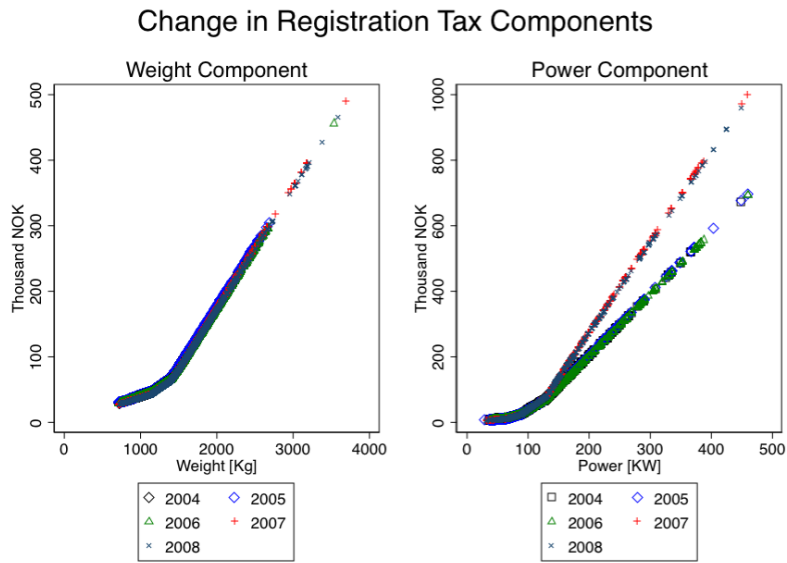


Figure 13: Scatter plot for the weight and power components of the registration tax for different years. The weight component remained almost constant, while the power component of the registration tax increased in 2007.

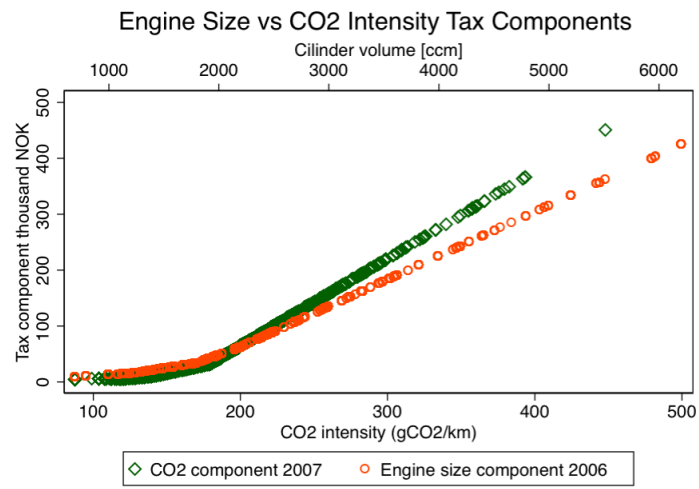


Figure 14: With the reform of 2007 the CO<sub>2</sub> intensity component of the registration tax substituted the engine component. The round scatter plot represents the amount in thousand NOK of the tax component calculated over the engine size in 2006. The diamond scatter plot represents the amount of tax paid for CO<sub>2</sub> intensity.

## B Comparison between control and treatment group

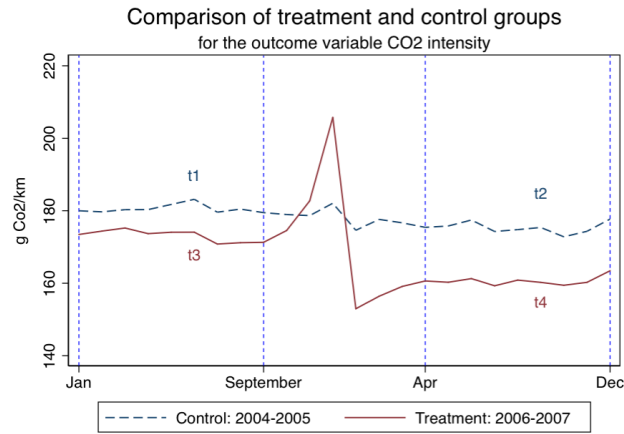


Figure 15: Comparison of treatment and control groups for the variable CO<sub>2</sub> intensity  $Control = t_1, t_2$  and  $Treatment = t_3, t_4$ , where  $t_1$  is April-December 2004,  $t_2$  is January-September 2005,  $t_3$  is January-September 2006 and  $t_4$  is April-December 2007.

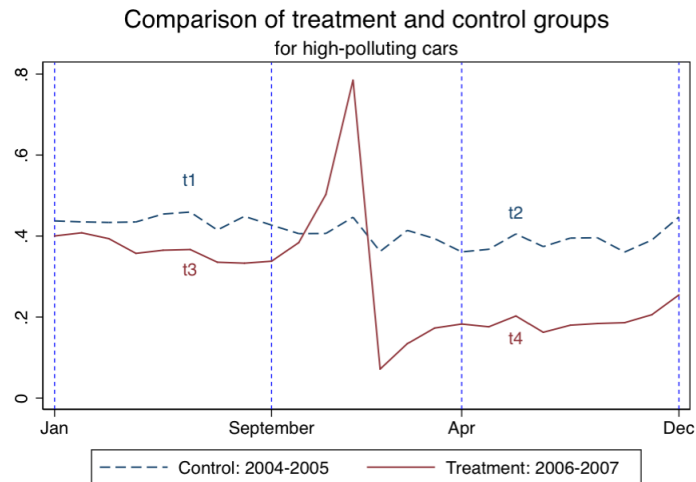


Figure 16: Comparison of treatment and control groups for cars emitting more than 180gCO<sub>2</sub>/km.  $Control = t_1, t_2$  and  $Treatment = t_3, t_4$ , where  $t_1$  is April-December 2004,  $t_2$  is January-September 2005,  $t_3$  is January-September 2006 and  $t_4$  is April-December 2007.

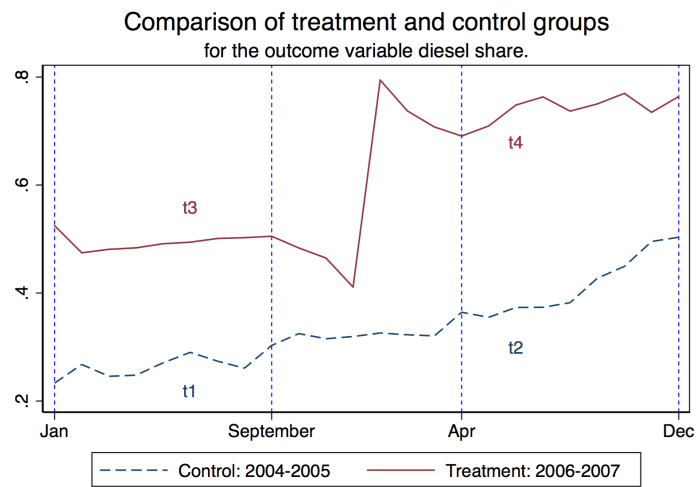


Figure 17: Comparison of treatment and control groups for diesel share. *Control* =  $t_1, t_2$  and *Treatment* =  $t_3, t_4$ , where  $t_1$  is April-December 2004,  $t_2$  is January-September 2005,  $t_3$  is January-September 2006 and  $t_4$  is April-December 2007.

## C Robustness check

To ensure the robustness of the estimation results presented in this work, the estimation for all outcome variables is carried out by postponing the after-treatment period. In the original specification the post-period consists of 9 months from April-December. Tables 8, 9, and 10 report the estimated coefficients when the post-treatment observations ( $t_2, t_4$ ) are taken from May-December, June-December, and July-December respectively. The results are robust to the specification used since the coefficients are stable.

Table 8: Robustness checks: Post-treatment May-Dec

	CO2 int.	High emis.	Diesel	N purchases	Weight	Power	Engine size
<b>Treatment eff</b>	-7.609*** (0.446)	-0.119*** (0.00658)	0.208*** (0.00848)	-0.420 (1.486)	5.527 (3.593)	0.609* (0.277)	46.48*** (4.265)
Group effect	-7.853*** (0.317)	-0.0822*** (0.00400)	0.177*** (0.00738)	-15.56*** (4.590)	49.00*** (2.483)	1.741*** (0.220)	13.34*** (3.802)
Time effect	-6.219*** (0.508)	-0.0627*** (0.00808)	0.118*** (0.00733)	-18.18*** (5.154)	22.91*** (4.459)	-0.490 (0.453)	-0.150 (6.048)
Gross Inc. (Mun.)	0.0414*** (0.0106)	0.000458*** (0.000118)	-0.000538*** (0.000147)	0.873*** (0.258)	0.197* (0.0917)	0.0341** (0.0114)	0.331* (0.159)
Diesel/Gas price ratio	-2.419 (2.551)	0.114*** (0.0333)	0.192*** (0.0284)	-22.47*** (4.387)	70.46*** (21.03)	9.514*** (2.358)	84.83* (35.93)
Fleet age (County)	-0.117 (0.518)	-0.00143 (0.00636)	0.0256*** (0.00558)	-2.972 (4.841)	7.663 (4.514)	0.112 (0.542)	9.331 (7.951)
Constant	171.8*** (6.533)	0.218** (0.0829)	-0.182 (0.0961)	-141.3* (70.48)	1130.7*** (55.94)	65.50*** (6.792)	1494.2*** (105.8)
Observations	12829	12829	22152	12829	12829	12829	12829
Adjusted $R^2$	0.529	0.449	0.335	0.151	0.259	0.126	0.113

Robust standard errors clustered on municipalities.

Table 9: Robustness checks: Post-treatment Jun-Dec

	CO2 int.	High emis.	Diesel	N purchases	Weight	Power	Engine size
<b>Treatment eff</b>	-7.536*** (0.452)	-0.123*** (0.00677)	0.201*** (0.00898)	-0.243 (1.483)	3.728 (3.629)	0.670* (0.281)	45.09*** (4.434)
Group effect	-7.860*** (0.324)	-0.0817*** (0.00410)	0.179*** (0.00736)	-15.65*** (4.572)	49.63*** (2.520)	1.807*** (0.227)	14.14*** (3.874)
Time effect	-6.323*** (0.533)	-0.0588*** (0.00870)	0.129*** (0.00780)	-18.53*** (5.183)	25.49*** (4.720)	-0.449 (0.479)	2.233 (6.498)
Gross Inc. (Mun.)	0.0445*** (0.0104)	0.000486*** (0.000116)	-0.000535*** (0.000154)	0.873*** (0.257)	0.222* (0.0896)	0.0370*** (0.0110)	0.371* (0.156)
Diesel/Gas price ratio	-3.401 (2.339)	0.0882* (0.0361)	0.0959*** (0.0271)	-19.67*** (3.930)	39.55* (19.90)	6.177** (1.995)	43.20 (31.57)
Fleet age (County)	-0.0784 (0.513)	-0.000957 (0.00630)	0.0258*** (0.00570)	-2.864 (4.770)	8.063 (4.435)	0.155 (0.530)	9.941 (7.769)
Constant	171.3*** (6.691)	0.228** (0.0847)	-0.0996 (0.0963)	-145.0* (69.12)	1146.5*** (56.47)	67.17*** (6.760)	1513.2*** (104.6)
Observations	12059	12059	20820	12059	12059	12059	12059
Adjusted $R^2$	0.520	0.434	0.337	0.151	0.267	0.140	0.119

Robust standard errors clustered on municipalities.

Table 10: Robustness checks: Post-treatment Jul-Dec

	CO2 int.	High emis.	Diesel	N purchases	Weight	Power	Engine size
<b>Treatment eff</b>	-7.211*** (0.441)	-0.123*** (0.00677)	0.192*** (0.00939)	0.0502 (1.481)	3.220 (3.743)	0.783** (0.297)	46.22*** (4.684)
Group effect	-7.886*** (0.327)	-0.0825*** (0.00418)	0.179*** (0.00736)	-15.77*** (4.573)	49.27*** (2.535)	1.805*** (0.232)	13.90*** (3.955)
Time effect	-6.778*** (0.553)	-0.0616*** (0.00897)	0.137*** (0.00818)	-18.97*** (5.232)	25.01*** (4.848)	-0.568 (0.510)	0.447 (6.936)
Gross Inc. (Mun.)	0.0469*** (0.0103)	0.000508*** (0.000122)	-0.000527*** (0.000152)	0.882*** (0.257)	0.238** (0.0907)	0.0392*** (0.0108)	0.407** (0.153)
Diesel/Gas price ratio	-3.433 (2.321)	0.104** (0.0373)	0.111*** (0.0255)	-20.90*** (3.884)	45.54* (19.23)	5.409** (1.873)	37.63 (30.82)
Fleet age (County)	-0.0494 (0.508)	-0.000735 (0.00635)	0.0263*** (0.00574)	-2.726 (4.684)	8.408 (4.487)	0.207 (0.526)	11.11 (7.713)
Constant	170.4*** (6.877)	0.206* (0.0894)	-0.120 (0.0973)	-147.6* (68.01)	1133.1*** (59.13)	66.67*** (6.913)	1495.9*** (105.3)
Observations	11280	11280	19487	11280	11280	11280	11280
Adjusted $R^2$	0.514	0.424	0.332	0.152	0.263	0.146	0.120

Robust standard errors clustered on municipalities.

## D Other Figures

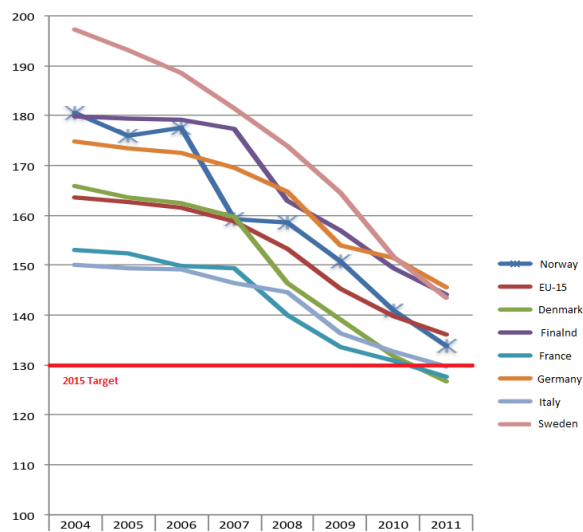


Figure 19: CO<sub>2</sub> intensity of new registered vehicles: a comparison between European countries. The reduction in CO<sub>2</sub> intensity for new registered vehicles in Norway is in line with the other European countries. Figure from OFV AS and Vista Analyse AS (Rapport 12/42) [http://www.regjeringen.no/pages/38231042/vista\\_rapport2012.pdf](http://www.regjeringen.no/pages/38231042/vista_rapport2012.pdf)

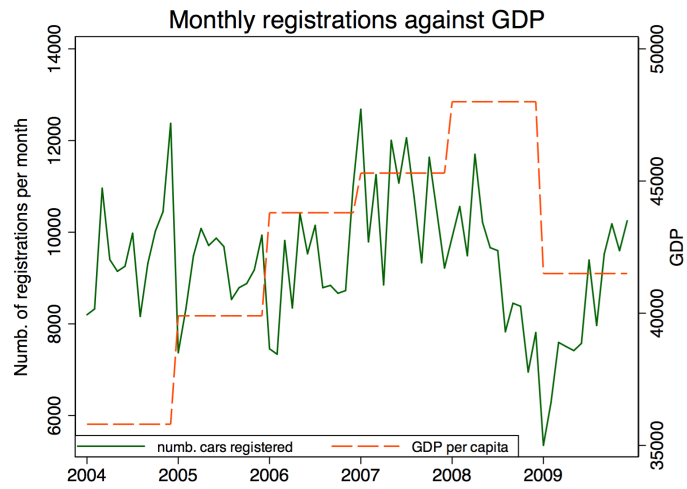


Figure 18: Graphic correlation between demand for private vehicles expressed in the number of new vehicles registered per month and GDP in Norway between 2004 and 2011. The drop in 2009 is probably due to the economic crisis, which had a mild effect on the Norwegian economy. The sales of vehicles recover strongly in 2009.

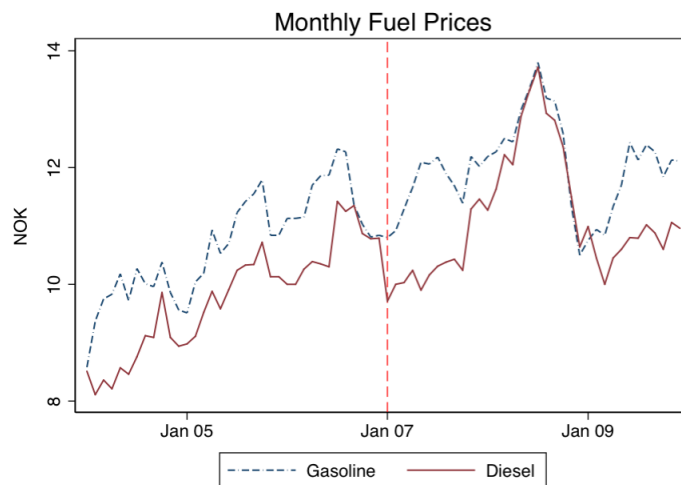


Figure 20: Monthly fuel prices including taxes.

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